


Interpretation of Isotherm Models for Adsorption of Ammonium onto Granular Activated Carbon

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Abstract: High amounts of ammonium (NH_4^+) discharged in receiving water can lead to eutrophication. The adsorption of NH_4^+ from synthetic solution onto granular activated carbon (GAC) was scrutinized with respect to initial solute concentration (10 mg L^{-1}), solution volume (0.2 L), adsorbent dosage (4 – 20 g), and contact time. Experimental data can be well described by the pseudo-second-order kinetic model ($R^2 > 0.994$) and Freundlich isotherm model ($R^2 = 0.936$), suggesting that chemisorption and multilayer adsorption occurred. Furthermore, this study explored the feasibility of using the Freundlich isotherm model to estimate the removal efficiency or required amount of adsorbent. The result findings indicated that GAC has a good potential to adsorb NH_4^+ from water and thus giving new insights into environmental engineering practices.

Keywords: adsorption; ammonium; eutrophication; isotherm model; granular activated carbon.

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1. Introduction

Nitrogenous matters such as ammonium, nitrite, and nitrate are an important element for animals and plants. However, the discharge of excessive amounts of these matters into surface water strongly accelerates eutrophication [1, 2]. Over enrichment of nitrogenous matters in the water, body expedites the plant growth and causing algal bloom. The water body may also lose its important functions and subsequently cause negative effects on the environment and human health [3, 4]. The effects of the release of nitrogen to surface water have led to legislations such as those by the United States Environmental Protection Agency (USEPA) and the European Union (EU) [5]. The USEPA permits the effluent limit of nitrogen must be less than 10 mg L^{-1} , while the EU only allows 15 mg L^{-1} of nitrogen for 10,000 – 100,000 population equivalents.

The excessive amounts of nitrogen can be treated with plenty of methods such as adsorption technique, biological nitrogen process, chemical process, and nanotechnology

application [6-9]. The biological process to treat wastewater can be highly variable due to operational difficulties, while the nanotechnology application and chemical treatment entail high chemical costs [10, 11]. From all these removal methods, adsorption is considered as a promising technique that could be employed in the removal of nitrogen from the water, given its simple operation and low-cost of the adsorbent materials [1, 12]. Nowadays, the use of GAC as an economic adsorbent for treating water has received extensive attention because of its potential in environmental and agronomic applications [13]. GAC can adsorb various pollutants due to high porous structure and high specific surface area. GAC has been used for the removal of specific chemicals such as potassium and phosphorus, which ultimately can be converted to fertilizer [14]. Moreover, GAC could also be used to remove nitrogen from water [15, 16].

Considering that the presence of algae in the water, with its chemical formula of $(\text{CH}_2\text{O})_{106} (\text{NH}_3)_{16} (\text{H}_3\text{PO}_4)$, resulting in eutrophication consists of ammonium (NH_4^+) [17], eutrophication can be controlled effectively by the removal of NH_4^+ . Hence, the aim of this study was to examine the feasibility of utilizing GAC for NH_4^+ removal from synthetic solutions with respect to adsorbent dosage (m), solute concentration (C_i), and contact time (t). The Freundlich and Langmuir isotherm models were employed to understand the mechanism of adsorption. Furthermore, this work aimed to elucidate how basic adsorption isotherm (i.e., Freundlich isotherm model) combined with material balance can be used to predict the value of m or removal efficiency (E) according to a given set of initial conditions. This work was performed to have an enabling methodology for the development of a computational model that can be used to make engineering predictions with quantified confidence. Besides that, it could be ideally applied to industrial applications.

2. Materials and Methods

2.1. Adsorption kinetics and isotherms.

2.1.1. Kinetic adsorption models.

Pseudo-first-order (PFO) and pseudo-second-order (PSO) kinetic models were used for describing the adsorption mechanism (Table 1). PFO equation can be expressed as given in Eq. (1) [15]:

$$\ln(q_e - q_t) = \ln(q_e) - k_1 t_i \quad (1)$$

PSO equation can be expressed as [16]:

$$\frac{t_i}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t_i}{q_e} \quad (2)$$

Eq. (1) can be used to determine the value of k_1 if the value of q_e has been verified through an experiment when its equilibrium adsorption is reached. The adsorption kinetic obeys a PFO model when the curve of plotting $\ln(q_e - q_t)$ against t_i gives a linear relationship depending on the adsorbate in aqueous solution [16].

A plot of t_i/q_t against t_i of the Eq. (2) should give a linear relationship with $1/q_e$ as slope and $1/k_2 q_e^2$ intercept at the vertical axis. The value of k_2 is a constant depending on the adsorbate with influence another parameter when the experimental adsorption kinetics data obey the PSO model [16].

Table 1. List of kinetic and isotherm adsorption models.

Models	Linear form	Plot	Parameters
<u>Kinetic</u>			
PFO	$\ln(q_e - q_t) = \ln(q_e) - k_1 t_i$	$\ln(q_e - q_t)$ vs. t_i	q_e k_1
PSO	$\frac{t_i}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t_i}{q_e}$	$\frac{t_i}{q_t}$ vs. t_i	q_e k_2
<u>Isotherm</u>			
Freundlich	$\ln q_e = \ln K_F + \frac{1}{n} \ln C_e$	$\ln q_e$ vs. $\ln C_e$	K_F n
Langmuir	$\frac{1}{q_e} = \frac{1}{K_L q_{max} C_e} + \frac{1}{q_{max}}$	$\frac{1}{q_e}$ vs. $\frac{1}{C_e}$	q_{max} K_L
<u>Nomenclature</u>			
k_1	PFO constant (min^{-1})		
k_2	PFO constant ($\text{g mg}^{-1} \text{min}^{-1}$)		
K_F	Freundlich constant (mg g^{-1})		
K_L	adsorption energy coefficient (L mg^{-1})		
n	heterogeneity factor (dimensionless)		
C_e	concentration of the adsorbate at equilibrium (mg L^{-1})		
q_e	adsorption capacity at equilibrium (mg g^{-1})		
q_t	adsorption capacity at time t (mg g^{-1})		
q_{max}	maximum adsorption capacity (mg g^{-1})		
t_i	adsorption time (min)		

2.1.2. Isotherm adsorption models.

The Freundlich and Langmuir equations are the equations commonly used to describe adsorption isotherms (Table 1). The Freundlich model describes that the heterogeneous surface of the adsorbates is formed on the surface of adsorbent with multilayer sorption of different energies of adsorption [18]. The Langmuir model assumes that the adsorbent is being saturated when the monolayer adsorbate coverage of adsorbent is attached with a homogenous surface without interactions between the adsorbed molecules [19]. The Freundlich equation can be written as [20]:

$$\ln q_e = \ln K_F + \frac{1}{n} \ln C_e \tag{3}$$

The Langmuir equation can be expressed as [16]:

$$\frac{1}{q_e} = \frac{1}{K_L q_{max} C_e} + \frac{1}{q_{max}} \tag{4}$$

The Freundlich isotherm suggests that a plot $\ln q_e$ against $\ln C_e$ of the Eq. (3) should give a straight-line intercept at K_F with $1/n$ as the slope. The adsorption coefficient K_F may indicate the affinity of adsorbate-adsorbent. The exponent n is related to the energetic heterogeneity of the adsorbent surface and determines either the favorable or unfavorable curve [21]. A plot of $1/q_e$ versus $1/C_e$ of Eq. (4) should give a straight line with $1/K_L q_{max}$ and $1/q_{max}$ as the slope and the intercept, respectively.

2.1.3. Freundlich isotherm model development.

The Freundlich isotherm can be expressed as [22]:

$$q_e = K_F C_e^{1/n} \tag{5}$$

The value of C_e and q_e can be calculated for a series of different conditions. Then $\ln q_e$ can be plotted as a function of $\ln C_e$. The adsorption isotherm obeys a Freundlich model when the curve of plotting $\ln q_e$ against $\ln C_e$ of Eq. (3) gives a straight line. Therefore, the values of $1/n$ and K_F can be determined from the slope and the intercept of Eq. (3).

The q_e in Eq. (5) can also be written as [22]:

$$q_e = \left(\frac{V C_e^S}{m} \right) \tag{6}$$

The isotherm equation was combined with a material balance [22]. In a batch study, the mass balance expression should be obeyed by the following expression:

$$C_i = C_e + C_e^S \tag{7}$$

where C_i and C_e^S are the solute concentration at the initial and the equilibrium, respectively. The q_e in Eq. (6) can also be written by the following equation [22]:

$$q_e = \left(\frac{V C_e^S}{m} \right) = \frac{V(C_i - C_e)}{m} \tag{8}$$

The Freundlich isotherm was rearranged with mass balance. Substituting Eq. (8) into Eq. (5) gives

$$\frac{K_F m}{V} C_e^{1/n} + C_e - C_i = 0 \tag{9}$$

in which C_e is a one-argument variable as follows. Even though Eq. (9) is not an explicit one and should be solved by using the numerical analysis method, C_e is a function of C_i and m for a specific V with parameters K_F and $1/n$. Therefore, with Eq. (10), $E = E(C_i, m)$.

Furthermore, E can be expressed as:

$$E = \frac{C_i - C_e}{C_i} \tag{10}$$

Substituting Eqs. (8) and (10) into Eq. (5), giving:

$$m = \frac{V}{K_F} \frac{E}{(1-E)^{1/n}} C_i^{(1-1/n)} \tag{11}$$

where $m = m(C_i, E)$ for a fixed V as shown in (Figure 1). Note that, Eq. (11) can be used to calculate the value of m for the desired value of E .

Therefore, when predicted m and E are obtained, adsorption q predicted can be estimated by substituting Eq. (10) into Eq. (8), giving:

$$q = \frac{E C_i V}{m} \tag{12}$$

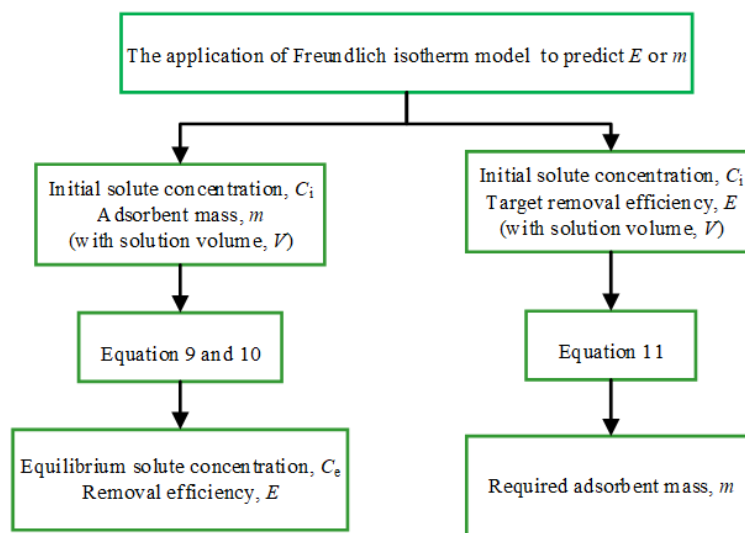


Figure 1. Diagram for the application of the Freundlich isotherm model for the estimation of E and m .

2.1.4. Freundlich isotherm parameter evaluation and model calculation.

The experimental data for NH_4^+ adsorption was fitted to the Freundlich model in Eq. (3) to evaluate isotherm parameters. The Newton-Raphson method [22] was applied to solve Eq. (9). The application tool for users was created using Excel to predict the value of E or m .

3. Results and Discussion

3.1. Adsorbents.

The commercial GAC was used as an adsorbent in this study. The GAC was purchased from Nikom Global Marketing (M) Sdn. Bhd., Selangor, Malaysia, and the raw material required to produce commercial GAC was coconut shell charcoal. Table 2 shows the characteristics of the GAC. The sample was washed several times with tap water. Next, the cleaned sample was oven-dried for 48 h at 30°C. The sample was crushed and then sieved to 0.60 mm.

Table 2. Main characteristics of the granular activated carbon.

Characteristic	Unit	Value
Origin (raw material)		Coconut shell
Bulk density	g cm ⁻³	0.52 – 0.54
Moisture contents	%	5
Hardness	%	97 – 98
Ash content	%	5
BET surface area	m ² g ⁻¹	1100
pH		9 – 10

3.2. Synthetic solutions.

Synthetic solution (NH₄⁺, 10 mg L⁻¹) was prepared by dissolving ammonium chloride (NH₄Cl) (analytic grade) into deionized water.

3.3. Analytical methods.

NH₄⁺ was determined using the Nessler method (HACH DR 6000, Spectrophotometer) [23]. Instrumental analysis was used to identify the characteristics of the GAC. In this work, the GAC samples were previously sputter-coated with gold (Sputter Coater, Model SC7620, Quorum Technologies, UK). A scanning electron microscope (SEM) (Model TM3000, Hitachi, Japan) was used to characterize the surface morphology of the unmodified GAC. The mineral phases in the GAC can be identified using a Bruker D8 advance high-resolution X-Ray Diffractometer (XRD). The chemical composition (in %) of GAC was analyzed using Energy Dispersive X-Ray Fluorescence (EDXRF) Spectrometer (Rigaku, Japan). The functional groups of the GAC were investigated using Fourier Transform Infrared (FTIR) Spectroscopy (IRTracer-100, Shimadzu, Japan). The surface area of GAC was performed by a multiple-point method according to the Brunauer, Emmett, and Teller (BET) theory, using a surfer analyzer (Surface Analyzer, Thermo Scientific Technologies, Italy).

3.4. Boehm titration.

The Boehm titration method was applied to determine the surface functional groups of GAC [24], where 0.3 g of GAC was added to 15 mL of sodium hydroxide (NaOH) (0.1 M), sodium carbonate (Na₂CO₃) (0.05 M), sodium hydrogen carbonate (NaHCO₃) (0.1 M) or hydrochloric acid (HCl) (0.1 M) solution in the flask, separately. The blanks and the sample solutions were shaken at a speed of 100 rpm and 25°C for 48 h. Then, 5 mL aliquots of each sample were filtered, and the excess of acid and bases was back-titrated with NaOH (0.1 M) and HCl (0.05 M), respectively. The number of basic surface functional groups was determined on the assumptions that HCl neutralized basic groups while the number of acidic surfaces functional groups were identified on the assumptions that NaHCO₃ neutralized only carboxylic

groups, Na₂CO₃ neutralized lactonic and carboxylic groups, and NaOH neutralized phenolic, lactonic and carboxylic groups [24].

3.5. Batch experiments.

The adsorption isotherm and kinetics were determined through batch experiments. The kinetic experiments were performed by adding 4, 12, and 20 g of the adsorbent, and the isotherm experiments were performed by adding 4, 8, 12, 16, and 20 g of the adsorbent into a different flask containing 0.2 L of 10 mg L⁻¹ synthetic solution. Each sample solution was shaken at 160 rpm, and then the concentrations of NH₄⁺ in each flask were identified at certain time intervals. The sample solutions were centrifuged, and the HACH DR 6000 UV–Vis Spectrophotometer was used to evaluate supernatant concentrations of NH₄⁺ present in each Erlenmeyer flask. The adsorption mechanism was determined by using the PFO, PSO, Langmuir, and Freundlich models. Each batch adsorption experiment was conducted twice, and the data obtained are the average values. The adsorption capacity (q) and the E were calculated using Eq. (8) and Eq. (10), respectively.

3.6. Physicochemical characteristics of GAC.

The surface functional groups of GAC were identified according to the Boehm titration method (Figure 2). The GAC consists of total acidic functional group (TAFG) and total basic functional group (TBFG) with the amounts of 1.35 and 1.03 mmol g⁻¹, respectively (Figure 2). The TAFG comprises carboxylic, lactonic, and phenolic groups with the amounts of 0.70, 0.50, and 0.15 mmol g⁻¹, respectively. The TAFG (1.35 mmol g⁻¹) of higher than the TBFG (1.03 mmol g⁻¹) was evaluated and could have more ability to adsorb more NH₄⁺ ions from the solution. In addition, the studied GAC has a BET surface area of 1100 m² g⁻¹ (Table 2).

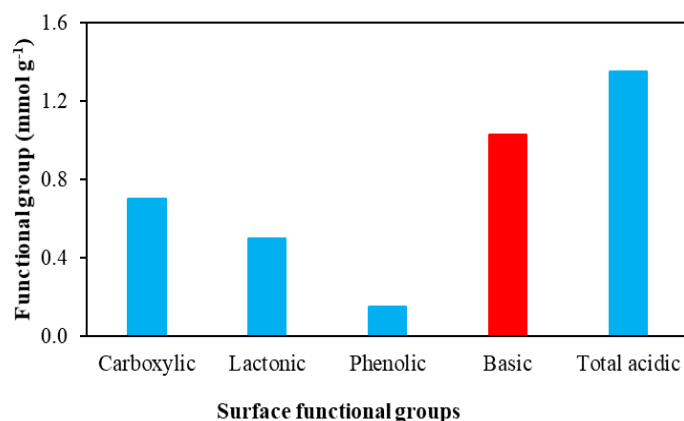


Figure 2. The surface functional groups of GAC.

The images from the SEM micrograph with 500, 2500, and 5000 times magnification were used to identify the surface morphology of the GAC (Figure 3). The surface feature of GAC has a coarse porous surface with irregular pores (Figure 3a). The images from the SEM micrograph with 2500 and 5000 times magnification has a rough texture with various pore size at the surface of the GAC (Figure 3b and Figure 3c).

The XRD pattern of the GAC is illustrated in Figure 4. The broad peaks 2θ of 26° and 2θ of 43° denote the graphitic structure [25]. The sharp peak at 2θ of 76° represents the quartz (SiO₂).

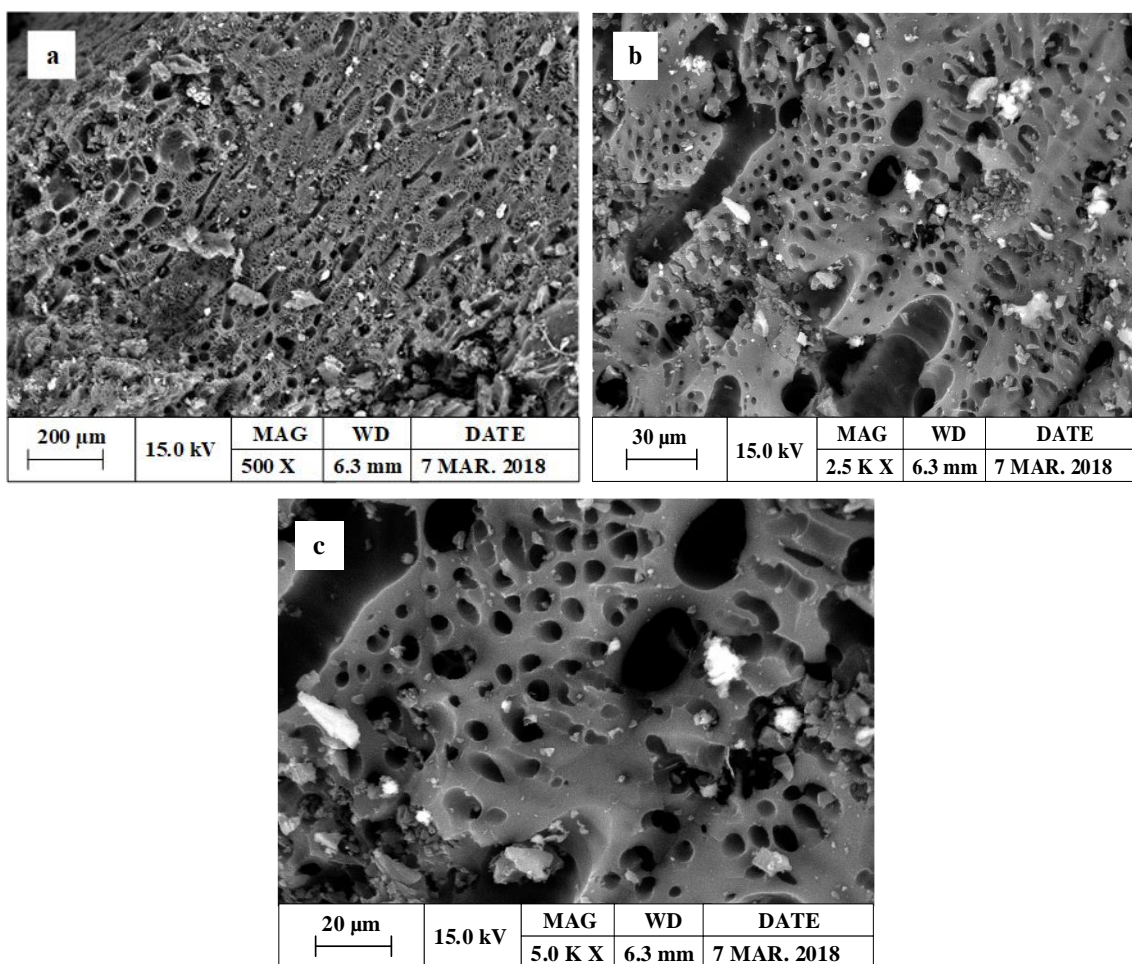


Figure 3. The images of GAC: (a) SEM images of 500× magnification (b) 2500× magnification (c) 5000× magnification.

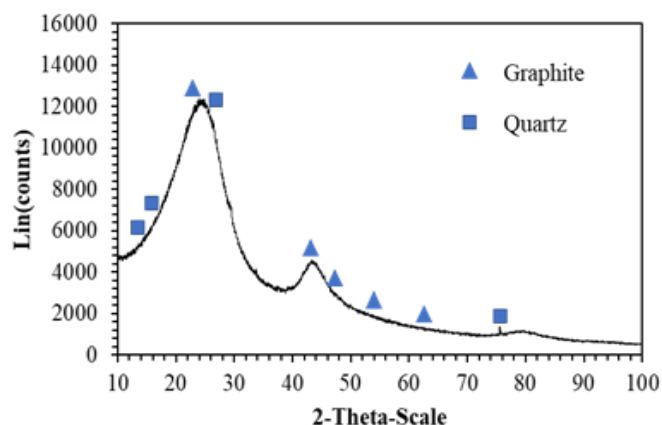


Figure 4. X-ray diffraction patterns of GAC.

The FTIR spectra analysis was carried out over the range from 600 to 4000 cm^{-1} (Figure 5). The bands at 1,645; 1,430; and 1,030 cm^{-1} indicative of C=O stretch; C–O stretch; and C–OH stretch [25]. After the adsorption of NH_4^+ , the band at 1,062 cm^{-1} corresponding to C–O stretching was detected and would be caused by the adsorption of NH_4^+ attached on the surface of GAC [27] while the band at 2,359 cm^{-1} could be assigned as N–H group [28, 29]. The vibration at a wavelength range of 3,400 to 3,700 cm^{-1} indicates the presence of O–H groups [30].

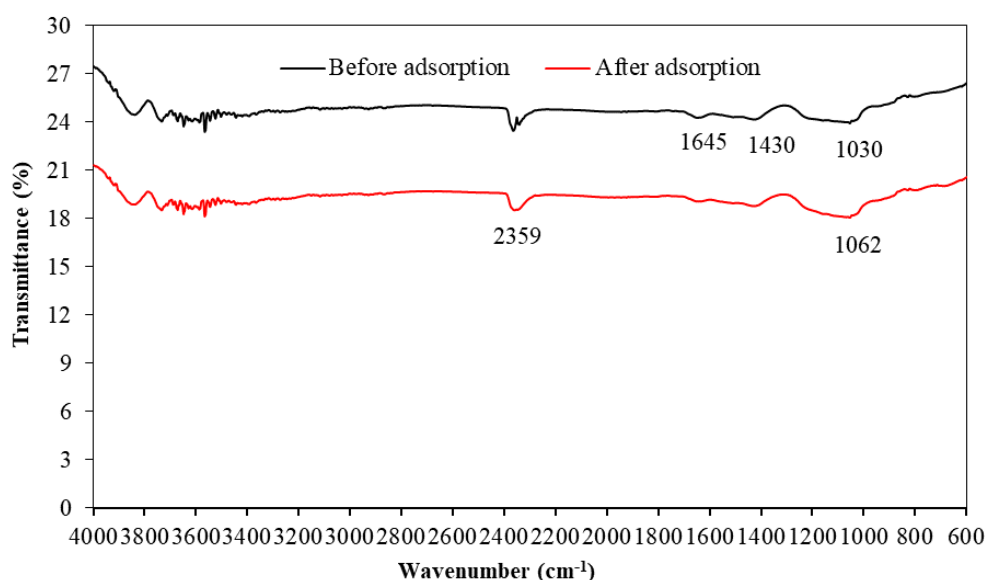


Figure 5. The FTIR spectra of GAC of before and after NH_4^+ adsorption.

3.7. Adsorption of NH_4^+ from a synthetic solution onto the GAC.

Figure 6 shows the variations of E pursuant to t for the removal of NH_4^+ using GAC. The use of GAC to adsorb NH_4^+ could be favorable because the acidic functional groups of carboxylic, phenolic, and hydroxyl presented at the surface of GAC having negative charges could be more favorable to react with NH_4^+ ions from synthetic solution to form the $\text{COO}^-(\text{NH}_4)^+$ complexes [31]. The efficiency of NH_4^+ removal can reach approximately 20.1, 50.1, and 70.4% for the removal of NH_4^+ after a contact time of 120 h with the amounts of GAC used to run the experiments were 4, 12, and 20 g, respectively. The E value to adsorb NH_4^+ onto GAC rapidly increases over a time period of 24 h and then slowly increases to reach equilibrium (Figure 6). The rapid adsorption may be due to the availability of a high number of free active sites of the acidic functional groups present on the surface of the GAC that may have a high capability to adsorb NH_4^+ ions from synthetic solution [32]. Subsequently, the rate of NH_4^+ adsorption is slow, which could be due to the abundance of active sites on the GAC that have been covered by NH_4^+ ; the adsorption equilibrium can be then achieved after 90 h (Figure 6).

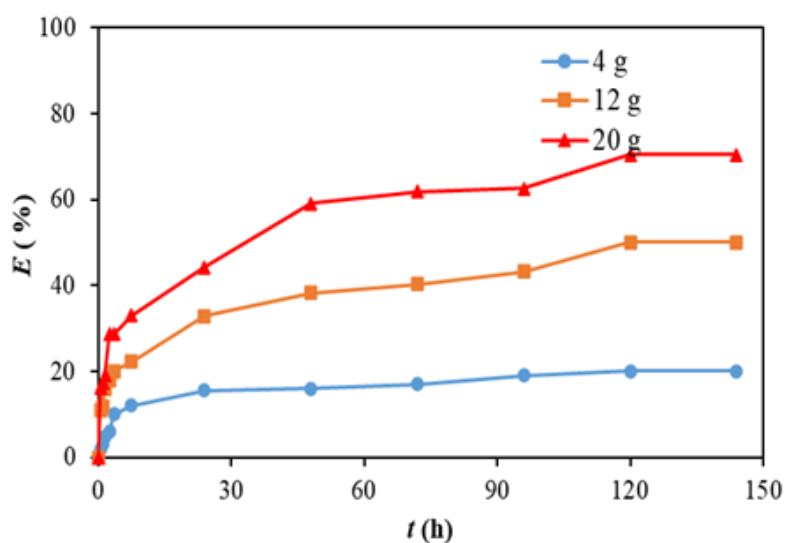


Figure 6. The efficiency of NH_4^+ removal from synthetic solution.

The plots of E pursuant m and q pursuant m for the adsorption of NH_4^+ from a synthetic solution onto GAC are shown in Figure 7. When the value of m increasing from 4 to 20 g, the value of E gradually increases from 20.1 to 70.4%, but q gradually decreases from 0.101 to 0.070 mg g^{-1} . More adsorbent used in a batch experiment could have more active sites available to attract NH_4^+ from a synthetic solution, and thus the E increases [33]. The q value decreases with an increasing amount of the GAC because the use of more GAC to adsorb NH_4^+ from synthetic solution could have more unoccupied active sites, and thus the q reduces [34-36].

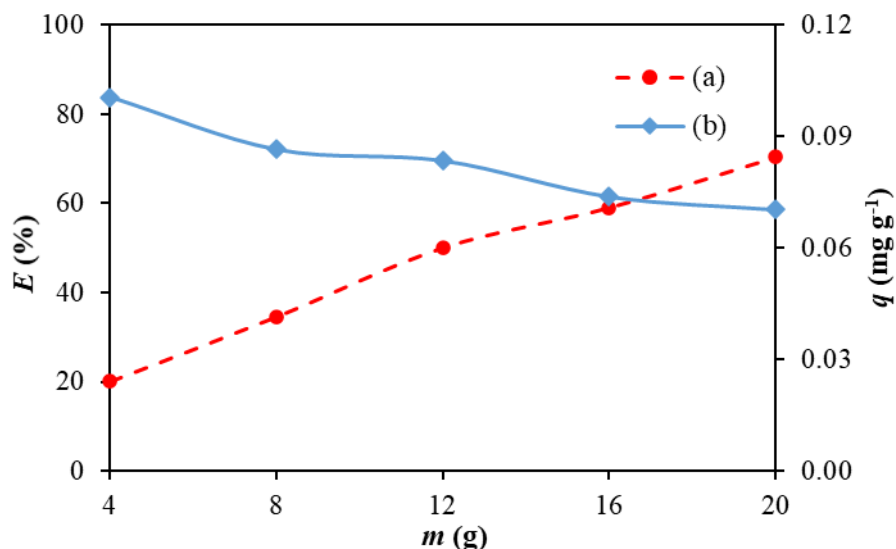


Figure 7. Relationship of: (a) E versus m and (b) q versus m for the NH_4^+ adsorption onto GAC.

3.8. Adsorption kinetics of NH_4^+ onto the GAC.

The F_e value can be calculated according to the following equation [37]. The most appropriate model, either PFO or PSO should have the smallest error function (F_e) value and the highest correlation coefficient (R^2) value.

$$F_e = \sqrt{\left(\frac{1}{n-p}\right) \sum_i^n (q_{t(\text{exp})} - q_{t(\text{theo})})^2} \quad (13)$$

where p and n are the numbers of kinetic parameters and measurements, respectively, $q_{t(\text{exp})}$ is the experimental q value (mg g^{-1}), and $q_{t(\text{theo})}$ is the theoretical q value (mg g^{-1}).

The linear regression analysis of the kinetic models for NH_4^+ is shown in Figure 8. The ($R^2 > 0.986$) for PSO model was higher than that ($R^2 > 0.908$) for PFO model, as shown in Table 3. This study verifies that PSO model could be more suitable for the adsorption kinetic of NH_4^+ onto GAC compared to PFO model due to the lower value of F_e and the higher value of R^2 have been evaluated (Table 3). According to the results of this study, the adsorption between GAC and NH_4^+ can be categorized as chemical adsorption because the adsorption process involving valency forces through sharing or exchange of electrons between the acidic functional groups and NH_4^+ ions [13, 38]. The value of k^2 increases from 0.025 to 0.028 and to 0.036 $\text{g mg}^{-1} \text{min}^{-1}$ with an increasing amount of the GAC from 4 to 12 and to 20 g, meaning that the value of k^2 is positively correlated with the value of m and thus, the rate of NH_4^+ adsorption onto GAC can be escalated by increasing the GAC dosage [16, 39].

Table 3. The kinetic parameters for PFO and PSO models.

Sample	Amount (g)	PFO model			F_e	q_e (exp) (mg g ⁻¹)
		q_e (theo) (mg g ⁻¹)	k_1 (min ⁻¹)	R^2		
Synthetic solution	4	0.068	0.0004	0.908	0.037	0.101
	12	0.056	0.0003	0.954	0.031	0.084
	20	0.047	0.0004	0.939	0.025	0.070
Sample	Amount (g)	PSO model			F_e	q_e (exp) (mg g ⁻¹)
		q_e (theo) (mg g ⁻¹)	k_2 (g mg ⁻¹ min ⁻¹)	R^2		
Synthetic solution	4	0.103	0.025	0.994	0.007	0.101
	12	0.083	0.028	0.986	0.019	0.084
	20	0.071	0.036	0.992	0.007	0.070

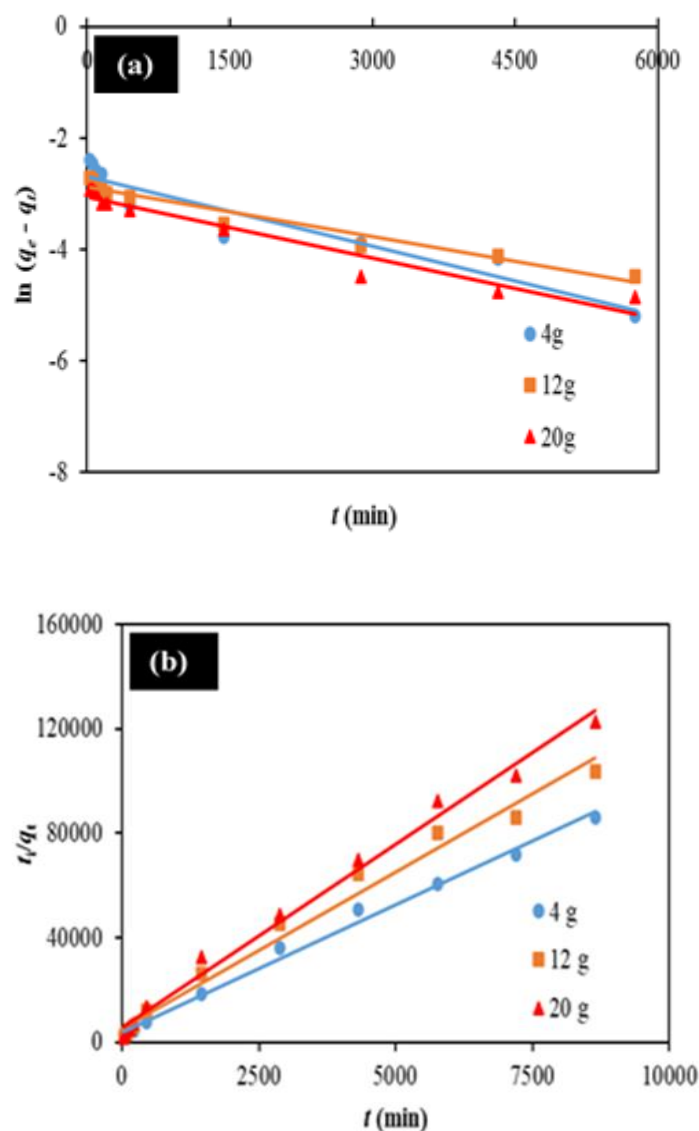


Figure 8. Experimental data of NH_4^+ adsorption onto GAC fitted to linear forms (a) PFO model and (b) PSO model.

3.9. Adsorption isotherms of NH_4^+ onto the GAC.

The Freundlich model (Figure 9a) and Langmuir model (Figure 9b) were used to analyzing the isotherms adsorption data. Table 4 shows the values of the isotherm parameters. This study found that the experimental data were best described by the Freundlich isotherm model ($R^2 = 0.9362$) compared to the Langmuir isotherm model ($R^2 = 0.9023$) as depicted in

Table 4. The adsorption of NH_4^+ occurred on the heterogeneous site of GAC progression with multilayer adsorption. The surfaces of the GAC are heterogeneous, and the sorption of NH_4^+ onto GAC occurs in the form of multilayers [6, 40]. The n value of 2.80 was verified (Table 4). The n value greater than one indicates favorable adsorption; active sites with the highest binding energies would be used first for less heterogeneous surfaces and then pursued by weaker sites for more heterogeneous surfaces [41-43].

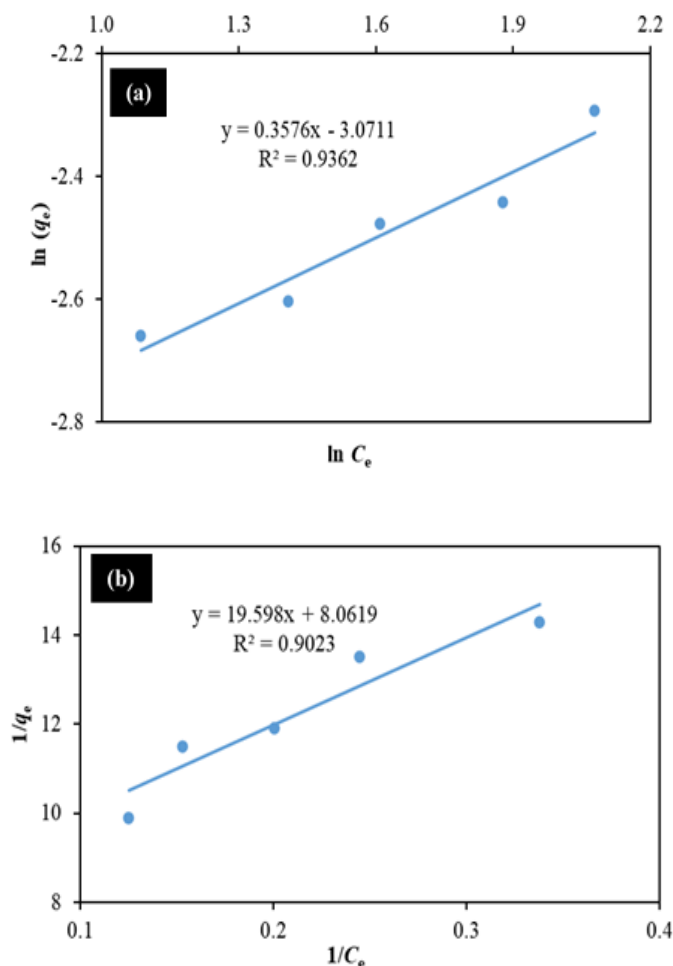


Figure 9. Experimental data of NH_4^+ adsorption onto GAC fitted to (a) Freundlich model and (b) Langmuir model.

Table 4. The isotherm parameters for Freundlich and Langmuir models.

Sample	Freundlich model			Langmuir model		
	n	K_F (mg g^{-1})	R^2	q_{max} (mg g^{-1})	K_L (L mg^{-1})	R^2
Synthetic solution	2.80	0.046	0.9362	0.124	0.411	0.9023

3.10. Application of Freundlich isotherm to estimate E or q .

The Freundlich and Langmuir isotherm models in Eq. (3) and Eq. (4) were used to analyze the isotherms adsorption data, which is depicted in Figure 9. The results show that the Freundlich model (Figure 9a) could be more suitable to explain the experimental adsorption data compared to the Langmuir model (Figure 9b). The isotherm parameters K_F and n for the Freundlich isotherm model are listed in Table 4.

The value of E for any set of experimental conditions can be estimated when the Freundlich parameters K_F and n were identified. Eq. (9) with Eq. (10) were used to obtain the

value of E predicted as a function of the initial concentration of NH_4^+ solution with different amount of the adsorbent from 4 to 20 g. The predicted data were compared to the experimental data for validation (Table 5 and 6). When $C_i = 10 \text{ mg L}^{-1}$ for the synthetic solution and $m = 4, 12, \text{ and } 20 \text{ g}$ for GAC in 0.2 L, E were predicted to be 19.4, 49.3, and 68.9% (Table 5). The experimental data values were 20.1, 50.1, and 70.4%, and thus, the prediction errors were 3.6, 1.6, and 2.1%, respectively. These findings show that when constants (i.e., K_F and n) in Eq. (3) were evaluated with an experimental set under specific conditions, the E in Eq. (10) can be predicted without significant error by using the model.

Furthermore, model verification was conducted to determine the required amount of adsorbent (Table 6) by using Eq. (11). In addition, when predicted m and E were obtained, q can be estimated by using Eq. (12). When $C_i = 10 \text{ mg L}^{-1}$ for the synthetic solution and desired $E = 21.0, 52.8, \text{ and } 72.5\%$, the required m was predicted to be around 4.4, 13.2, and 22.0 g L^{-1} and thus, the prediction errors were 9.1, 9.9, and 9.8%, respectively. The adsorption capacities were predicted to be 0.10, 0.08, and 0.07 mg g^{-1} with an increasing amount of the GAC from 4.4 to 13.2 to 22.0 g, and the prediction errors were 4.3, 4.1, and 6.2%, respectively. These findings show that the m required to eliminate pollutant/solute at the desired E can be predicted.

Table 5. Comparison of E experimental and E predicted.

Sample	Initial conditions		Experimental E (%)	Predicted E (%)	Error (%)
	C_i (mg L^{-1})	m (g)			
Synthetic solution	10	4	20.1	19.4	3.6
	10	12	50.1	49.3	1.6
	10	20	70.4	68.9	2.1

Table 6. Comparison of m experimental and m predicted.

Sample	Initial conditions		Experimental m (g)	Predicted m (g)	Error (%)	Predicted q (mg g^{-1})	Error (%)
	C_i (mg L^{-1})	Desired E (%)					
Synthetic solution	10	21.0	4.0	4.4	9.1	0.10	4.3
	10	52.8	12.0	13.2	9.9	0.08	4.1
	10	72.5	20.0	22.0	9.8	0.07	6.2

3.10.1. Practical application of the Freundlich isotherms.

The result of Figure 10 was generated by plotting E against m and q against m with the initial synthetic solution with a concentration of 10 mg L^{-1} . This figure illustrates that when desired $E = 72.5\%$ at $C_i = 10 \text{ mg L}^{-1}$, m value of 22 g should be supplied in 1 L solution and thus $q = 0.07 \text{ mg g}^{-1}$ (points A and B in Figure 10), meaning that 22 g of GAC should be used to eliminate 72.5% of NH_4^+ from 10 mg L^{-1} synthetic solution in a 1 L batch adsorption reactor.

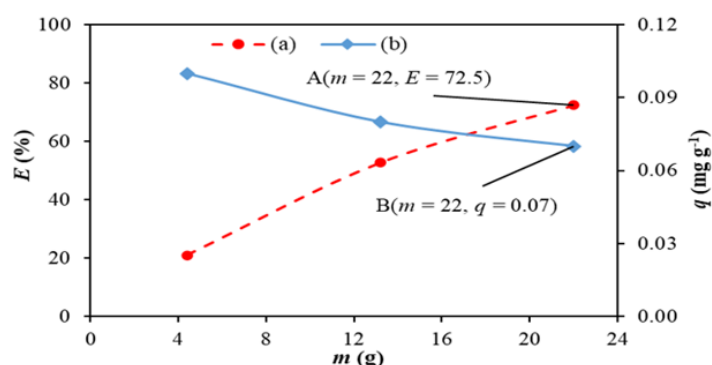


Figure 10. Relationship of: (a) E pursuant m and (b) q pursuant m for the adsorption of NH_4^+ from synthetic solution onto GAC.

4. Conclusions

In this research, the adsorption of NH_4^+ onto GAC from synthetic solution was best described by the Freundlich model, implying that the adsorption process would occur as a multilayer. The result findings can help to solve the excessive amounts of NH_4^+ problem by using GAC as adsorbent and thus potentially improve environmental quality. Moreover, the isotherm model with mass balance permits estimation of the m needed for the desired E at an initial solute concentration, indicating that this model is practical in the adsorption process used for pollutant removal from water.

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Conflicts of Interest

The authors declare no conflict of interest.

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